

BIO

by Ahmad Faizal

FILE	01.PUBLISHEDRJET_FEB2016_AHMAD_FAIZAL_FAK_KELAUTAN.PDF (231.99K)	WORD COUNT	5499
TIME SUBMITTED	22-APR-2019 02:27PM (UTC+0700)	CHARACTER COUNT	29592
SUBMISSION ID	1116795757		



Research Journal of
**Environmental
Toxicology**

ISSN 1819-3420



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Research Article

Bioavailability of Pb and Cu in Sediments of Vegetated Seagrass, *Enhalus acoroides*, from Spermonde Islands, Makassar, South Sulawesi, Indonesia

Shinta Werorilangi, Muh. Farid Samawi, Rastina, Akbar Tahir, Ahmad Faizal and Arniati Massinai

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Department of Marine Science, Faculty of Marine Science and Fisheries, Hasanuddin University, Tamalanrea Km. 10, Makassar 90245, Indonesia

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Abstract

Seagrass might impact metal bioavailability in sediments with physiological processes that occur at the root and rhizoma. Seagrass growth may increase bioavailability due to oxygen transport from leaves to the root systems. This study aims to analyze the concentrations of bioavailable Pb and Cu in sediments with and without seagrass. This study was done at two sites in Spermonde Islands, South Sulawesi, Indonesia. All metals analysis sediment parameters were conducted on dry, <63 µm grain size sediment samples. Metal speciation in sediment was determined using the Community Bureau of Reference (CBR) three steps sequential that extract exchangeable and acid soluble fraction, reducible and oxidisable fraction. The average concentrations of bioavailable fraction of Pb and Cu (fraction 1) was higher in vegetated sediments associated with *Enhalus acoroides* than in unvegetated sediments. Higher concentration of Cu in fraction 1 was also associated with higher concentration in *Enhalus* roots. This indicates that the presence of seagrass may increase the bioavailability of metals in sediments. Increased metal bioavailability in vegetated sediment will imply increase toxicity.

Key words: Speciation, metals, seagrass, sediment, Spermonde islands

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Received: December 16, 2015

Accepted: February 08, 2016

Published: February 15, 2016

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Citation: Shinta Werorilangi, Muh. Farid Samawi, Rastina, Akbar Tahir, Ahmad Faizal and Arniati Massinai, 2016. Bioavailability of Pb and Cu in sediments of vegetated seagrass, *Enhalus acoroides*, from Spermonde Islands, Makassar, South Sulawesi, Indonesia. Res. J. Environ. Toxicol., 10: 126-134.

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Corresponding Author: Shinta Werorilangi, Department of Marine Science, Faculty of Marine Science and Fisheries, Hasanuddin University, Tamalanrea Km. 10, Makassar 90245, Indonesia

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Competing Interest: The authors have declared that no competing interest exists.

Data Availability: All relevant data are within the paper and its supporting information files.

INTRODUCTION

Anthropogenic pollutants, such as metals can reduce the distribution and biomass of seagrass (Macinnis-Ng and Ralph, 2002). Bioavailability of metals in sediments is strongly influenced by processes that occur in the marine environment. This could affect the availability of metals to organisms that would affect further toxicity (Bernhard and Neff, 2001).

Metal bioavailability is influenced by sediment characteristics such as pH, redox oxidation potential, dissolved organic carbon, inorganic complexes, organic complexes and sediment particle size. Changes in redox potential in sediment will affect metal mobility. Study shows that an increase of oxygen content in anoxic sediments will increase metal released to surrounding environments and thus available to organisms (Clark *et al.*, 1998; Eggleton and Thomas, 2004; Kelderman and Osman, 2007). Dredging activities in sediment, bioturbation and seagrass physiological activities in root systems can influence oxidation state in sediments. Bioturbation by bottom organisms, oligochaetes and benthic bivalves, increases metal bioavailability by changing the redox potential through sediment resuspension (Peterson *et al.*, 1996; Ciutat and Boudou, 2003; Atkinson *et al.*, 2007).

Seagrass can absorb metals from the water column and sediment through the leaf and root-rhizome and distributed to various compartments in plants (Prange and Dennison, 2000; Macinnis-Ng and Ralph, 2002; Sanz-Lazaro *et al.*, 2012). In coastal waters where the dissolved oxygen is low (suboxic to anoxic), heavy metals will generally be strongly bound to sulphites in sediments and reduce its bioavailability for the organism. However, marine plants such as seagrass can oxidize sediments in the root zone, through the diffusion of oxygen from the leaves to the roots, causing the oxidation of sulphite-metal compounds that can release metals into the sediment which then increases the potential bioavailability to biota (Pulich, 1987; Weis and Weis, 2004).

Lead (Pb) is a toxic metal, found naturally in the waters and has no biological function for the organism (non-essential). Human activities on land (anthropogenic), especially the use of additives in gasoline, paint industries, battery, increasing the concentration of Pb in the waters. According to Neff (2002), the majority of Pb in waters originate from the atmosphere. Copper (Cu) is an essential metal for the metabolism of biota, but increasing concentrations above the standard minimum requirement will have deleterious effects on biota. The use of Cu a raw material in antifouling paints (TBT) is one of the main sources of these metals entering the sea (Srinivasan and Swain, 2007; Bao *et al.*, 2008). With the increasing development of

coastal areas of Makassar, anthropogenic inputs of this metal will also affect sediment and biota living on it.

This study describes differences in Pb and Cu bioavailability in sediment of seagrass and non seagrass areas off the coast of Makassar, Spermonde Islands, South Sulawesi. The result aims to increase understanding of seagrass oxidation impact on metal bioavailability in sediment. It also improves knowledge of metal contamination and toxicity in different sediment types. Information from this study will improve risk assessment method for metal toxicity in sediments.

MATERIALS AND METHODS

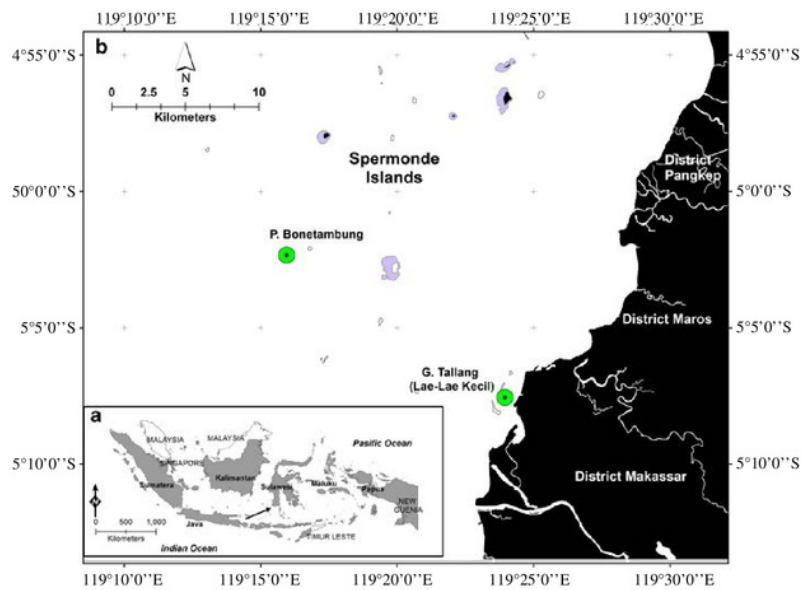
Study sites: This study was conducted in Bonetambung Island and Gusung Tallang (Lae Lae Caddi Island), which are part of The Spermonde Islands, near Makassar, South Sulawesi, from March-April, 2013. Gusung Tallang and Bonetambung island are located approximately 1 and 17 km from coast, respectively (Fig. 1). Seagrass beds in Gusung Tallang has monospecies of *Enhalus acoroides*, whereas, in Bonetambung Island has multi-species and dominated by *Enhalus acoroides*. Sediment and seagrass samples were taken from seagrass and non-seagrass areas at each island.

Sample collection

Sediment: There were three replicates of vegetated (seagrass) and non-vegetated (non seagrass) sediment in each island. Sediment was sampled from the vegetated and non vegetated areas from oxic part of sediment surface (1-3 cm depth) using van Veen grab sampler. Samples were specifically taken from the center of the grab to avoid possible contamination from metal parts on the grab and placed into 1 kg plastic bags in a cool box during transport to the laboratory and then frozen in a freezer at 4°C until processing.

Seagrass: Three shoots of *Enhalus acoroides* were randomly collected from the vegetated area in each island. Below ground part (root and rhizome) of the plants were taken for metal content analysis. Prior to the metal analysis, the below ground part were stored overnight at -20°C.

Metal and bioavailability analysis: Sediment and below ground part of seagrass were analyzed for lead (Pb) and copper (Cu). Before analysis, sediments were dried in room temperatures for 4 days and oven dried for 16 h at 80°C. Sediment samples were sieved in dry conditions to attain sediment particles of <63 µm (Loring and Rantala, 1992; Yuan *et al.*, 2004; Hendozko *et al.*, 2010).



3 Fig. 1: Study sites in Spermonde Islands, South Sulawesi, Indonesia

Table 1: Three steps sequential procedures determining metals bioavailability in sediments

Steps	Fractions	45 46 47 48 49 50 51 52 53 54 55 56 57 58 59 60 61 62 63 64 65 66 67 68 69 70 71 72 73 74 75 76 77 78 79 80 81 82 83 84 85 86 87 88 89 90 91 92 93 94 95 96 97 98 99 100 101 102 103 104 105 106 107 108 109 110 111 112 113 114 115 116 117 118 119 120 121 122 123 124 125 126 127 128 129 130 131 132 133 134 135 136 137 138 139 140 141 142 143 144 145 146 147 148 149 150 151 152 153 154 155 156 157 158 159 160 161 162 163 164 165 166 167 168 169 170 171 172 173 174 175 176 177 178 179 180 181 182 183 184 185 186 187 188 189 190 191 192 193 194 195 196 197 198 199 200 201 202 203 204 205 206 207 208 209 210 211 212 213 214 215 216 217 218 219 220 221 222 223 224 225 226 227 228 229 230 231 232 233 234 235 236 237 238 239 240 241 242 243 244 245 246 247 248 249 250 251 252 253 254 255 256 257 258 259 260 261 262 263 264 265 266 267 268 269 270 271 272 273 274 275 276 277 278 279 280 281 282 283 284 285 286 287 288 289 290 291 292 293 294 295 296 297 298 299 300 301 302 303 304 305 306 307 308 309 310 311 312 313 314 315 316 317 318 319 320 321 322 323 324 325 326 327 328 329 330 331 332 333 334 335 336 337 338 339 340 341 342 343 344 345 346 347 348 349 350 351 352 353 354 355 356 357 358 359 360 361 362 363 364 365 366 367 368 369 370 371 372 373 374 375 376 377 378 379 380 381 382 383 384 385 386 387 388 389 390 391 392 393 394 395 396 397 398 399 400 401 402 403 404 405 406 407 408 409 410 411 412 413 414 415 416 417 418 419 420 421 422 423 424 425 426 427 428 429 430 431 432 433 434 435 436 437 438 439 440 441 442 443 444 445 446 447 448 449 450 451 452 453 454 455 456 457 458 459 460 461 462 463 464 465 466 467 468 469 470 471 472 473 474 475 476 477 478 479 480 481 482 483 484 485 486 487 488 489 490 491 492 493 494 495 496 497 498 499 500 501 502 503 504 505 506 507 508 509 510 511 512 513 514 515 516 517 518 519 520 521 522 523 524 525 526 527 528 529 530 531 532 533 534 535 536 537 538 539 540 541 542 543 544 545 546 547 548 549 550 551 552 553 554 555 556 557 558 559 560 561 562 563 564 565 566 567 568 569 570 571 572 573 574 575 576 577 578 579 580 581 582 583 584 585 586 587 588 589 590 591 592 593 594 595 596 597 598 599 600 601 602 603 604 605 606 607 608 609 610 611 612 613 614 615 616 617 618 619 620 621 622 623 624 625 626 627 628 629 630 631 632 633 634 635 636 637 638 639 640 641 642 643 644 645 646 647 648 649 650 651 652 653 654 655 656 657 658 659 660 661 662 663 664 665 666 667 668 669 670 671 672 673 674 675 676 677 678 679 680 681 682 683 684 685 686 687 688 689 690 691 692 693 694 695 696 697 698 699 700 701 702 703 704 705 706 707 708 709 710 711 712 713 714 715 716 717 718 719 720 721 722 723 724 725 726 727 728 729 730 731 732 733 734 735 736 737 738 739 740 741 742 743 744 745 746 747 748 749 750 751 752 753 754 755 756 757 758 759 760 761 762 763 764 765 766 767 768 769 770 771 772 773 774 775 776 777 778 779 780 781 782 783 784 785 786 787 788 789 790 791 792 793 794 795 796 797 798 799 800 801 802 803 804 805 806 807 808 809 810 811 812 813 814 815 816 817 818 819 820 821 822 823 824 825 826 827 828 829 830 831 832 833 834 835 836 837 838 839 840 841 842 843 844 845 846 847 848 849 850 851 852 853 854 855 856 857 858 859 860 861 862 863 864 865 866 867 868 869 870 871 872 873 874 875 876 877 878 879 880 881 882 883 884 885 886 887 888 889 890 891 892 893 894 895 896 897 898 899 900 901 902 903 904 905 906 907 908 909 910 911 912 913 914 915 916 917 918 919 920 921 922 923 924 925 926 927 928 929 930 931 932 933 934 935 936 937 938 939 940 941 942 943 944 945 946 947 948 949 950 951 952 953 954 955 956 957 958 959 960 961 962 963 964 965 966 967 968 969 970 971 972 973 974 975 976 977 978 979 980 981 982 983 984 985 986 987 988 989 990 991 992 993 994 995 996 997 998 999 1000	Procedures
1	Acid soluble	exchangeable ions and carbonates	40 mL CH ₃ COOH added on 1 g dry sediments (<63 μm), 16 h shaker
2	Reducible	Fe-Mn oxides	40 mL NH ₂ OH.HCl (pH 2 with HNO ₃), 16 h shaker
3	Oxidizable	Organic compounds and sulfites	10 mL H ₂ O ₂ , 1 h digestion, 1 h water bath, added 10 mL H ₂ O ₂ , 1 h water bath, added 50 mL CH ₃ COONH ₄ (pH 2 with HNO ₃)

The bioavailable fractions of Pb and Cu in sediments was determined by the Community Bureau of Reference (BCR) Three-Steps Sequential Method (Ure *et al.*, 1993). The extractions procedures were summarized in Table 1. All reagents were of extra pure quality and all lab wares used were either new or thoroughly cleaned with 10% HNO₃ for 24 h before utilization.

Prior to metal extractions on seagrasses, below ground parts (roots and rhizomes) were cut in pieces and oven dry for 48 h in 60°C. Total concentration of Pb, Cu and Fe was analysed as a sediment character) in sediments and total concentrations of Pb and Cu at the below ground part of the seagrasses were extracted by dry destruction method using HNO₃ (nitric acid) and HClO₄ (per chloric acid) (USEPA., 1994). Analysis of metals concentrations were performed on Atomic Absorption Spectrophotometer (AAS) Shimadzu AA-7000.

Sediment character analysis: Sediment samples in the oxic layer of fine grain sediment (<63 μm) were also analysed for: Total Organic Carbon (TOC) content by Walkley-Black titration method (ASTM., 2000); sulfur content spectrophotometrically

using HNO₃ dan HClO₄ (Tabatabai and Bremner, 1970); CaCO₃ by titrimetric method (Allison and Moodi, 1965); sediment texture by hydrometer method, type and sediment grain size based on Wentworth Scale using dry sieving method (Boggs, 2006). Sediment redox potential was measured using Hanna Instrument (HI 8314) with ionode probe, Australia (IJ64).

Statistical analysis: Independent sample T-student test was performed to determine differences in the average concentration of metals in sediments and seagrasses from both study areas. Correlation analysis amongst metal concentrations in each fraction and sediment geochemical parameters was performed using Pearson Correlation. All statistical analysis were computed using SPSS version 16 and the graphs were performed with Microsoft Excel for Windows.

RESULTS

Sediment characteristics: Characteristics and geochemical properties of the sediments in the two sites, Gusung Tallang (Lae Lae Kecil) and Bonetambung Island, are presented in

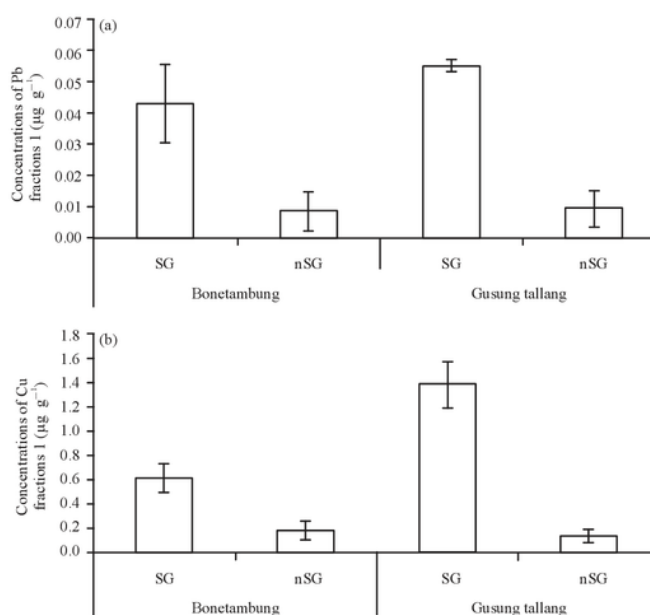


Fig. 2(a-b): Average concentrations of Pb and Cu in bioavailable fractions (fraction 1) in seagrass (SG) and non-seagrass (nSG) sediments from Bonetambung Island and Gusung Tallang (Mean ± SE, n = 3)

Table 2: Sediment characteristics in Bonetambung Island dan Gusung Tallang

	Bonetambung		Gusung Tallang	
	SG	nSG	SG	nSG
Sand (%)	92.85 ± 0.17	92.95 ± 0.47	92.62 ± 0.28	92.63 ± 0.49
Mud (%)	7.15 ± 0.17	7.05 ± 0.47	7.38 ± 0.28	7.37 ± 0.49
CaCO ₃ (%)	3.61 ± 0.87	1.20 ± 0.25	6.52 ± 2.93	7.54 ± 1.22
TOC (%)	1.53 ± 0.07	1.55 ± 0.16	1.72 ± 0.11	1.88 ± 0.20
Sulfur (%)	1.82 ± 0.26	1.96 ± 0.06	1.13 ± 0.26	2.01 ± 0.50
Total Fe (µg g ⁻¹)	6.95 ± 1.47	6.66 ± 1.09	70.07 ± 9.63	51.34 ± 1.89
Redox Potential (mV)	51.33 ± 19.29	91.90 ± 24.61	-115.33 ± 44.20	-52.37 ± 22.81
GS < 0.63 mm	0.89 ± 0.15	0.67 ± 0.11	0.22 ± 0.03	2.66 ± 0.76

*SG: seagrass and nSG: non-seagrass (mean ± SE, n = 3)

Table 2. As can be seen, non-seagrass sediment of Gusung Tallang has slightly higher fine grain particles, but in general, sediment texture conditions on both locations consisted of >90% sand and only about 7% which is mud (silt+clay). Total organic carbon and sulfur content in both sites has also similar values.

Of all sediment parameters analysed, CaCO₃, sulfur content and redox condition are much different comparing both study areas. Calcium carbonate content is almost double and sulfur content is ten times higher in Gusung Tallang's than in Bonetambung Island. Sediment redox potential shows the more reduced condition in Gusung Tallang; whereas, the more oxidized condition in Bonetambung Island.

Metals in bioavailable fractions from seagrass and non-seagrass sediment: Only metals in fraction 1 are presented in this study, because fraction 1 (acid soluble fraction) has the most mobility and thus the most related to the bioavailable concentration. Average concentrations of Pb and Cu in fraction 1 are shown in Fig. 2.

Mean concentrations of ⁴³Pb in fraction 1 in both study areas show significantly higher in seagrass than in non-seagrass sediment (p < 0.05). The Pb concentrations in fraction 1 is also significantly higher in seagrass than non-seagrass sediment (p < 0.05) in Gusung Tallang. Although, is not statistically significant, Concentrations of Pb in fraction 1 of seagrass sediment from Bonetambung Island is also higher than those of non-seagrass.

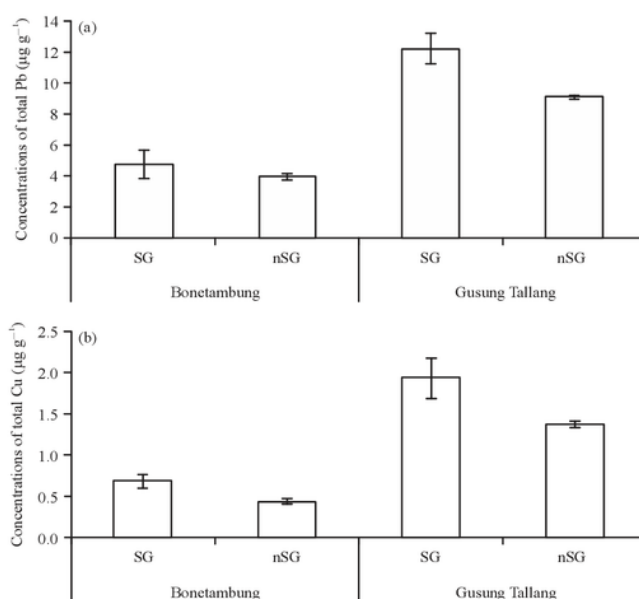


Fig. 3(a-b): Average concentrations of total Pb and Cu in seagrass (SG) and non-seagrass (nSG) sediments from Bonetambung Island and Gusung Tallang (Mean ± SE, n = 3)

Only average concentration of Cu in fraction 1 from seagrass area in sediment from Gusung Tallang that is significantly higher comparing to those from Bonetambung Island ($p < 0.05$).

Seagrass and total sediment metal concentrations in relation to bioavailable metal concentrations in sediments:

Average concentrations of total Pb and Cu in sediment on both study areas are presented in Fig. 3. Concentrations of total Cu in sediment from Bonetambung Island and total Pb from Gusung Tallang are significantly higher in seagrass area than non-seagrass area ($p < 0.05$). Total concentrations of Pb and Cu in each sediment area are almost three times higher in Gusung Tallang than in Bonetambung Island ($p < 0.01$).

Average Pb and Cu concentrations in are shown in Fig. 4. Concentrations ($\mu\text{g g}^{-1}$) of Pb in *Enhalus* roots from Bonetambung Island and Gusung Tallang ranged from 1.90-3.68 (avg 2.52 ± 0.58) and 2.78-5.66 (avg 4.15 ± 0.83), respectively. Whereas, Cu concentrations ($\mu\text{g g}^{-1}$) 0.40-0.86 (avg 0.56 ± 0.15) and 2.78-5.67 (avg 1.45 ± 0.32), respectively. Although, there are no significant different between mean concentrations of Pb and Cu in seagrass roots from both Bonetambung Island and Gusung Tallang ($p > 0.05$), but from the graph in Fig. 4, average concentrations of both metals are higher in seagrass roots from Gusung Tallang than from Bonetambung Island.

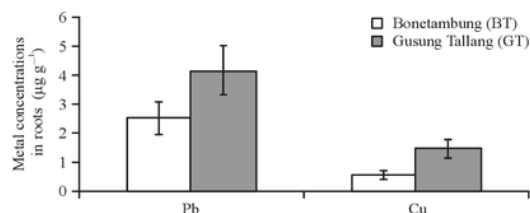


Fig. 4: Average concentrations of Pb and Cu in *Enhalus acoroides* roots (Mean ± SE, n = 3)

Relation between bioavailable fraction with total metal concentrations in sediment and seagrass roots:

Bioavailable fractions of metals consist of all three fractions (fraction 1, 2 and 3) from speciation study. The most bioavailable and the most toxic form is fraction 1 because of high solubility and thus easily enter the organisms body. The other fraction (2 and 3) which are also the non-resistant, can be easily removed from sediment matrix and solubilized, they become potentially toxic to organisms depending of physical and chemical parameters of such as redox potential changes, oxygen content and pH changes (Ramirez *et al.*, 2005).

In this study, the pearson correlation analysis (Table 3) were performed between all metal fractions (1, 2 and 3) and total concentration in sediment and in *Enhalus acoroides* roots. There is a strong positive correlation between Cu

concentrations in *Enhalus* roots and its total concentration in sediment ($r = 0.868$, $p < 0.05$), fractions 1 and 3 ($r = 0.848$ and $r = 0.876$, respectively, $p < 0.05$). There are no correlation between Pb in seagrass root with all metals analysed.

DISCUSSION

Sediment characteristics: While the sampling sites were selected to contrast terrigenous and biogenic sediment, it is not clear from the sediment analysis if the sediments should be classified as terrigenous or biogenic. Based on Al-Rousan *et al.* (2006), sediment type at the two sites is terrigenous because both have CaCO_3 content of $< 10\%$ (Table 2). However, based on Fe content at both locations, sediment types in Gusung Tallang is more of terrigenous types, whereas Bonetambung Island is more biogenic sediment types. Badr *et al.* (2009) stated that iron enrichment in coastal areas constituted a 27% of ferromagnesium from terrigenous material. Also, according to Chen *et al.* (1996), the iron content in biogenic sediments are generally very low due to the concentration of Fe on the surface of the sea water is very small. Gusung Tallang has higher content of iron (Fe) due to close proximity to the mainland and therefore sources of pollution from Makassar City.

These results contradicted to Erfteimeijer and Middelburg (1993) who found a high content of CaCO_3 ($> 90\%$) in the sediment of Barranglompo Island, the island adjacent to Bonetambung Island; while in Gusung Tallang, a location close to the mainland (Makassar) contains about 10% of CaCO_3 .

The CaCO_3 and Fe content of sediments may influence bioavailability of metals in sediment. According to John and Leventhal (1995), the most bioavailable fractions (dissolved fractions) consists of carbonate complexes which tend to increase with increasing pH and metals will be released with decreasing pH. The medium mobility is metals bound to Fe and Mn oxides in surface particulates matters and or sediments. Metals released is governed by redox potential in sediment, i.e. with higher redox potential (high oxidation states) tend to increase metal solubility and in reduced conditions, sulfate is reduced and metal will bound tightly to sulfide.

The higher average redox potential was measured in Bonetambung Island (71.63 mV) and according to Colman and Holland (2000), this was a suboxic condition of sediments; whereas, Gusung Tallang's sediment has a transitional redox potential from suboxic to anoxic condition (-83.85 mV). Redox potential condition may influence metals mobility in sediments. In oxic condition metals bound in sediment materials may be released to the water column and become available to organisms.

Metal in bioavailable fractions from seagrass and non-seagrass sediments: It is clear from the speciation study that most bioavailable fractions (fraction 1) of Pb and Cu are higher in sediments with seagrass. This indicated that seagrass plant may influence metal bioavailability in sediments. According to Weis and Weis (2004), aquatic plants can oxidize sediments around the roots through oxygen which is transported from leaves to the roots. Mangrove *Avicennia* was found to oxidize below ground part of the plants which can lead to sulphites reduction and increasing metal concentrations in the exchangeable fraction (fractions 1).

Study conducted by Doyle and Otte (1997) also found higher metal concentrations in vegetated areas than non-vegetated areas, especially in the area around root plants. In the seagrass plants, transport oxygen from the leaves to the roots is used for respiration and nutrient uptake, due to slightly stretched membrane on the roots some oxygen will leach to sediment and will lead to occurrence of oxidation processes near the seagrass roots (Swartz *et al.*, 2004). The oxidation processes can further lead to the release of sulphites-bound metals from sediments.

Higher Pb and Cu concentrations in the most bioavailable fraction 1 in Gusung Tallang compare to Bonetambung Island is most probably due to the close proximity to the mainland where anthropogenic wastes maybe transported from. According to Yap *et al.* (2002) and Yap and Wong (2011), metals in the non-resistant fractions (fraction 1, 2 and 3) are an indication that the source of metals are most likely due to anthropogenic input rather than natural input.

Although, sediment Pb and Cu concentration found in this study were below Sediment Quality Guidelines (SQGs) generated by NOAA's (Long *et al.*, 1995), i.e., Effects Range-Low (ERL) and Effects Range-Medium (ERM) concentration for Pb (46.7 and 218 $\mu\text{g g}^{-1}$, respectively) and Cu (34 and 270 $\mu\text{g g}^{-1}$, respectively). Similarly, Ambo-Rappe (2014) also found Pb and Cu concentration in sediment from seagrass area in Gusung Tallang and Bonetambung Island were still below the guideline for protection of aquatic biota. However, the bioavailability results in this study indicated possible risks of metal toxicity on organisms associated with seagrass.

Seagrass and total sediment metal concentrations in relation to bioavailable metal concentrations in sediments: Concentrations of Pb and Cu in *Enhalus* roots reflect the total metals concentrations in sediments, i.e., both total metals concentrations in sediment are higher in Gusung Tallang than Bonetambung Island (Fig. 3). Higher metals in Gusung Tallang is predicted because it is located near Makassar city, which is may contribute more metals to the areas.

Table 3: Pearson Correlation among metals in *Enhalus acoroides* and sediments

Concentrations	Seagrass root	
	Pb	Cu
Sediment (total)	0.658	0.868*
Fraction 1	-0.047	0.848*
Fraction 2	0.221	-0.051
Fraction 3	0.771	0.876*

*Correlation is significant at the 0.05 level (2-tailed), n = 3

Seagrass can absorb metals from the water column and sediment (Ambo-Rappe *et al.*, 2007). Absorption of metals from the water column can be performed through the plant leaves and from sediments through the roots and rhizomes. Once the metal is absorbed by seagrass, translocation may occur from the top to the bottom of the seagrass or otherwise. Essential metals such as Cu and Mn, will generally accumulate in the leaves because of metabolic needs; whereas the non-essential metals such as Pb and Cd will be accumulated more in the below ground of seagrasses, i.e. the roots and rhizomes (Prange and Dennison, 2000; Wasserman and Wasserman, 2002). In line with these study, the average total Pb concentration in the rooting section was higher than Cu in both study sites.

The positive significant Pearson Correlation analysis in Table 3 may indicate accumulation of Cu in seagrass sourced from metals in sediments and also closely associated with Cu in fraction 1 and fraction 3. Higher proportions of Cu in fraction 1 may increase its bioavailability and therefore could increase the concentration of Cu in the seagrasses, thus its toxicity to the plant. However, low average concentrations of Cu in seagrass roots compared to Pb further indicates the possibility of translocation of Cu from the below ground to the above ground of seagrasses, i.e., leaf, where necessary for metabolic processes of plants.

A significant positive correlation between the concentration of Cu in seagrass root with Cu in fractions 3 showed that the higher Cu bound to organic materials, sulfite and carbonate, the higher metal accumulate in the roots of seagrasses. With the changes in pH and redox potential from anoxic to oxic conditions, may release metals that bind strongly to the fraction 3 to become more easily absorbed by seagrasses. Changes in the oxidation condition may be due to the transfer of oxygen from the leaves to the roots. The possible higher metabolic rate (photosynthetic rate) of seagrass plants in Gusung Tallang than in Bonetambung Islands was indicated by higher Cu in fraction 1 in Gusung Tallang when compared to Bonetambung Island (Fig. 4).

Arifin *et al.* (2012) studied metal contamination in sediment of Indonesian waters from Jakarta Bay and Berau Delta, East Kalimantan. They found higher concentration of Pb in reducible fraction (fraction 2) in contaminated sediment

from Jakarta Bay; whereas, in Berau Delta, concentration of Pb mostly in residual fraction that is strongly bound to sediment matrix, thus not available for uptake by organisms.

The apparent absence of a significant correlation between the average concentration of Pb in seagrasses with metal in sediments and fractions may indicate that there were other factors affected the absorption of Pb to the seagrasses. Nevertheless, there was a fairly high correlation between Pb concentrations in seagrasses with the metal in sediments, this could indicate the source of Pb in seagrass roots systems derived from sediments.

CONCLUSION

This is the first study on bioavailability of metals on seagrass sediment. This study found the higher average concentrations of bioavailable fraction of Pb and Cu (fraction 1) are related to the vegetated sediments associated with *Enhalus acoroides*. Higher concentration of Cu in fraction 1 was also associated with higher concentration of Cu in *Enhalus* roots. This indicates that the presence of seagrass may increase the bioavailability of metals in sediments. Increased metal bioavailability in vegetated sediment will imply increase toxicity. Higher concentration of Pb and Cu at the root of seagrasses was related to the higher total metal concentration in sediments; suggested source of metals comes from sediments. Higher metals concentration in bioavailable fractions, in sediment bulk and also in *Enhalus* root from Gusung Tallang, especially Cu, further indicated that the source of Pb and Cu are most probably from anthropogenic input rather than natural input.

ACKNOWLEDGEMENTS

We would like to thank Indonesian Government for the funding provided by Indonesian Directorate General of Higher Education (DIKTI) through BOEN-Hasanuddin University 2013. Thanks are extended to the Laboratory of Chemical Oceanography, Faculty of Marine Science and Fisheries, Hasanuddin University and Isyanita for the space and lab works assistance, to Dr. Rohani Ambo-Rappe and Dr. Joanne Wilson for the reading and improving the manuscript prior to submission.

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